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# The state of the art review on photocatalytic Cr(VI) reduction over MOFs-based photocatalysts: From batch experiment to continuous operation

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### HIGHLIGHTS

- The Cr(VI) reduction in MOF-based photocatalysts in past 5 years were summarized.
- The continuous Cr(VI) reduction can be achieved via the immobilized MOFs.
- The prospective of Cr(VI) reduction in MOF-based photocatalysts is declared.

### GRAPHICAL ABSTRACT



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### ABSTRACT

This state of the art review presented the photocatalytic reduction from highly toxic Cr(VI) to lowly toxic Cr(III) with metal-organic frameworks (MOFs) and their composites. The construction of composites facilitated the transportation of the photo-induced charges to enhance the Cr(VI) reduction, in which the corresponding mechanisms were clarified by both experimental tests and DFT calculations. The immobilized MOFs onto some substrates accomplished continuous operations toward Cr(VI) reduction even under real solar light. As well, the environmental applications of the Cr(VI) reduction were analyzed, in which the influence factors toward the Cr (VI) reduction were clarified. This review reported that a big breakthrough was achieved from the batch experiment to the continuous operation for Cr(VI) reduction, in which MOFs demonstrated a bright prospective in the field of photocatalytic Cr(VI) reduction.

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### 1. Introduction

Some metal-organic frameworks (MOFs) including but not limited to UiO-66, MIL-125, MOF-5 are deemed as ideal photocatalyst candidates (Wang et al., 2014; Zeng et al., 2016), in which the organic ligands and metal-oxo clusters act as the light-absorbing antenna and isolated semiconductor quantum dots, respectively (Wang and Ho, 2016; Wang et al., 2020). The metal templates, organic linkers, channels or pores in the MOFs frameworks can be easily tuned due to the intrinsic modular property of the MOFs (Wang et al., 2020). Especially, the well-defined crystal structures of MOFs can be obtained from the single crystals via the X-ray diffraction technology, which can be used to clarify the structure-activity relationship (Wang et al., 2022; Zhao et al., 2021a). Therefore, the MOFs-based photocatalysts attracted increasing interests in various research fields like H<sub>2</sub> and O<sub>2</sub> production from water splitting (Ali et al., 2021; Cai et al., 2017), CO<sub>2</sub> reduction (Wang et al., 2015a), organic pollutants degradation (Yi et al., 2021; Zhao et al., 2022), antimicrobial (Li et al., 2021a) and Cr(VI) reduction (Gao et al., 2019; He et al., 2021; Wang et al., 2016).

Considering that the MOFs constructed from the hard metals (like  ${\rm Zr}^{4+}$ ,  ${\rm Ti}^{4+}$  and  ${\rm Fe}^{3+}$ ) and polycarboxylic acids (like terephthalic acid, trimesic acid and fumaric acid) can be stable under acidic solutions, they are often adopted as photocatalysts to accomplish photocatalytic hexavalent chromium  ${\rm Cr}({\rm VI})$  reduction (Wang et al., 2019). As well known, all  ${\rm Cr}({\rm VI})$  compounds are highly toxic considering their high water solubility and carcinogenic impact toward most organisms (Chen et al., 2021; Kaur et al., 2021), which can be reduced into  ${\rm Cr}({\rm III})$  to low the toxicity and to increase the precipitation (Wei et al., 2016; Yao et al., 2014; Zhang et al., 2015). Among the different  ${\rm Cr}({\rm VI})$  sequestration methods like chemical reduction, photocatalytic reduction displayed some advantages like convenient utilization of solar light, less or even no secondary pollution, as well as lower cost, which is preferred to be used for the removal of  ${\rm Cr}({\rm VI})$  pollutants in water (Li et al., 2018; Zhang et al., 2018); Zhang et al., 2013).

In 2016, we published a mini review paper to present the research progress of the Cr(VI) reduction with the help of MOFs photocatalysts (Wang et al., 2016), which attracted massive attentions from our worldwide counterparts. In the past five years, some significant research findings concerning the photocatalytic Cr(VI) reduction adopting MOFs and MOFs-based materials as photocatalysts emerged. Also, some review papers concerning Cr(VI) reduction were published, in which the corresponding research status and outlook were proposed as from different views (Kumar et al., 2021; Li et al., 2021; Zhao et al., 2019). In recent years, our research group focused on the photocatalytic Cr(VI) reduction over MOFs-based materials ranging from new pristine MOFs to different MOFs composites constructed from different secondary or ternary components. Especially, we proposed to accomplish continuous Cr(VI) reduction over the immobilized MOFs photocatalysts under artificial UV/visible light and even real solar light. Therefore, it was essential to present the updated review paper based on our corresponding research results for providing insightful summary and prospective, aiming to promote the large-scale applications in the real scenarios.

### 2. From individual MOFs to their composites

### 2.1. The new pristine MOFs as photocatalysts for Cr(VI) reduction

Inspired by the typical MOFs photocatalysts like NH<sub>2</sub>-MIL-125(Ti) (Wang et al., 2015b), NH<sub>2</sub>-UiO-66(Zr) (Shen et al., 2015; Shen et al., 2013a), NH<sub>2</sub>-MIL-88 B(Fe) (Shi et al., 2015), NH<sub>2</sub>-MIL-53(Fe) (Liang et al., 2015a), NH<sub>2</sub>-MIL-68(In) (Liang et al., 2015b) and NH<sub>2</sub>-MIL-101 (Fe) (Shi et al., 2015) for Cr(VI) reduction, some new approaches to enhance the photocatalytic Cr(VI) reduction were proposed. For example, Wang and coworkers fabricated (OH)<sub>2</sub>-UiO-66-X% via introducing 2,5-dihydroxyterephthalic acid to partially replace benzene-1,

4-dicarboxylic acid (Li et al., 2021b), due to that the introduction of –OH groups could harvest light with longer wavelength, narrow the band gap as well as enhance the separation of charge carriers (Xie et al., 2020). In this work, tartaric acid was adopted not only as electron transfer medium to produce COO· radicals for boosting the Cr(VI) reduction but also as hole capture agent to enhance the separation of the photo-yielded electron-hole pairs (Fig. 1). Besides the classic MOFs, some new MOFs photocatalysts like BUC-20 (Zhong et al., 2019), BUC-21 (Wang et al., 2017), BUC-66/67 (Yi et al., 2018), BUC-85/86 (Ma et al., 2020), NNU-36 (Zhao et al., 2017) and so on (Guo et al., 2020; Kaur et al., 2020; Xie et al., 2020; Zheng et al., 2019) were designed and synthesized for the purpose of photocatalytic Cr(VI) reduction.

Among these newly developed MOFs, BUC-21, a water stable 2D MOF constructed from cis-1.3-dibenzyl-2-imidazolidone-4.5-dicarboxylic acid, 4,4'-bipyridine and Zn2+, displayed the outstanding photocatalysis performance under the irradiation of UV light (Wang et al., 2017), due to that cis-1,3-dibenzyl-2-imidazolidone-4,5-dicarboxylic acid is an excellent light responsive organic linker. BUC-21 could accomplish 96% Cr(VI) reduction with initial K<sub>2</sub>Cr<sub>2</sub>O<sub>7</sub> concentration of 10.0 mg/L (pH = 2.0) within 30.0 min under UV light, which was superior to the P25's (a commercial TiO<sub>2</sub>) performance (only 39.0%) under the identical condition. Upon the irradiation of UV light, the photoinduced electrons were produced over the lowest unoccupied molecular orbital (LUMO) to reduce Cr(VI) into Cr(III). To enhance the Cr(VI) reduction via boosted separation of electron-hole pairs, the reactive red X-3B (X-3B) was selected as organic pollutant to consume the left holes over the highest occupied molecular orbital (HOMO) due to that BUC-21 can accomplish effectively photocatalytic degradation toward several organic dyes like methylene blue (MB), rhodamine B (RhB), methyl orange (MO) and X-3B under UV light. However, both Cr (VI) reduction and X-3B degradation activities decreased toward the matrix of Cr(VI) and X-3B with the identical conditions. In this system, the X-3B was primarily oxidized by  $\cdot O_2^-$  radicals rather than holes, in which the  $\cdot O_2^-$  radicals were formed via the reaction between the electrons and dissolved oxygen (DO). As the presence of both Cr(VI) and X-3B, the photo-induced electrons preferred to participate in the Cr(VI) reduction, which inhibited the formation of  $\cdot O_2^-$  radicals responsible to X-3B oxidation.

It was generally believed that the -NH2 groups in the ligand of the MOFs can accomplish visible light response and boosted separation of photo-induced electron-hole pair. For example, UiO-66 with band energy of 3.9 eV (HOMO: 3.35 eV and LUMO: 0.55 eV) can only be activated by UV light. However, the NH2-UiO-66 can be activated by visible light due to the narrower band energy of 2.52 eV (HOMO: 2.01 eV and LUMO: 0.51 eV). As well, the (OH)2-UiO-66 arose increasing interest due to that the introduced -OH groups can act as the intramolecular hole scavenger to enhance the quick separation of the photogenerated charge carrier (Li et al., 2021bLi et al., 2021b, 2021). However, the pristine (OH)<sub>2</sub>-UiO-66 suffered from the unregular morphology (Fig. 1a) (Xie et al., 2020). Our group proposed an approach to prepare (OH)2-UiO-66-X% (X represented the mole ratio of the introduced 2, 5-dihydroxyterephthalic acid) with mixed ligands of terephthalic acid and 2,5-dihydroxyterephthalic acid, in which the octahedral (OH)<sub>2</sub>-UiO-66-20% with the particle size ranging from 1300 nm to 1500 nm displayed outstanding photocatalytic Cr(VI) reduction under 369 nm UV light provided by low power 10 W LED light source (Fig. 1b). The (OH)2-UiO-66-20% displayed superior photocatalysis performance to the counterparts, in which the Cr(VI) with initial concentration of 10 mg/L can be completely reduced into Cr(III) at pH = 2.0 within 40.0 min. Tartaric acid (TA) as typical small molecular organic acid was adopted to further improve the Cr(VI) reduction performance. Without the addition of TA, the photo-induced electrons can reduce Cr(VI) via both direct electron transfer and via indirect electron transfer of forming  $\cdot$ O<sub>2</sub> radicals after reaction with dissolved oxygen (DO) (Fig. 1c). Beside the above-mentioned direct and indirect electron transfer mechanisms,

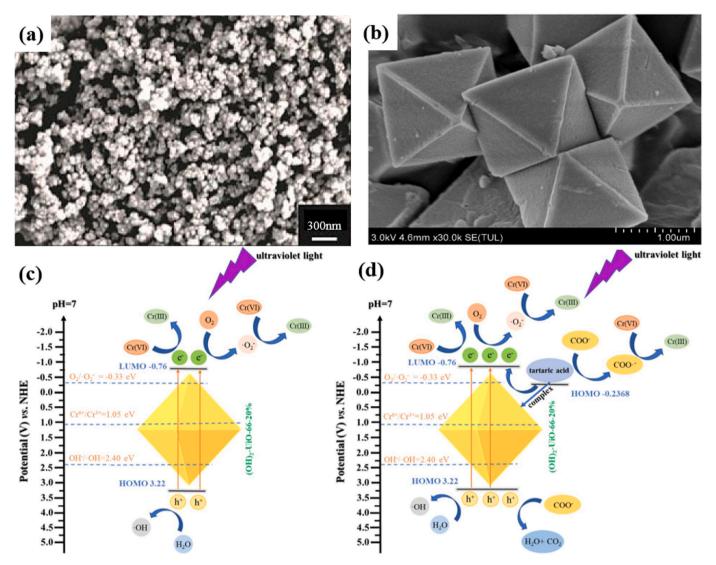


Fig. 1. SEM image of (a)  $(OH)_2$ -UiO-66 (Xie et al., 2020) and (b)  $(OH)_2$ -UiO-66-20% (Li et al., 2021b). The proposed mechanisms of Cr(VI) reduction in  $(OH)_2$ -UiO-66-20% (c) without the presence and (d) with the presence of tartaric acid (Li et al., 2021b).

the presence of TA can (i) consume the formed holes in HOMO to improve the electron-hole separation, (ii) react with (OH) $_2$ -UiO-66-20% photocatalyst to yield charge-transfer-complex (CTC) for accelerate electron transfer, and (iii) form COO- radicals to boost the Cr(VI) reduction (Fig. 1d). This work made it possible to prepare surface modified MOF photocatalysts, and to enhance the photocatalysis performance via the addition of suitable small molecular organic acid for practical applications.

### 2.2. The MOFs composites as photocatalysts for Cr(VI) reduction

Due to that the individual MOFs suffered from their poor electronic conductivity, rapid photo-induced charge carrier recombination, and limited light absorption utilization ability (Chen et al., 2020b; Li et al., 2021; Wang et al., 2014, 2016), it was a feasible and effective approach to combine MOFs with some highly conductive semiconductors or conductors like g-C<sub>3</sub>N<sub>4</sub> (Du et al., 2019; Gao et al., 2019; Yi et al., 2019), WO<sub>3</sub> (Wang et al., 2021b), TiO<sub>2</sub>-based semiconductors (Li et al., 2020b), bismuth-based semiconductors (Guo et al., 2020; Zhao et al., 2021b; Zhao et al., 2020a), Ag-based semiconductors (Zhou et al., 2020a; Zhou et al., 2019), N-K<sub>2</sub>Ti<sub>4</sub>O<sub>9</sub> (He et al., 2021), Cd<sub>0-5</sub>Zn<sub>0-5</sub>S (Wei et al., 2020) PTCDA (3,4,9,10-perylenetetracarboxylic dianhydride) (Wei et al., 2021), and PANI (polyaniline) (Chen et al., 2020a) with good

light-harvesting activity. It was believed that the introduction of some secondary or even ternary component to fabricate the composites with MOFs can also improve their stability (Ding et al., 2019), which can widen their potential applications (Wu et al., 2021).

Up to now, type-II and Z-scheme heterojunctions, as two mainstream heterojunctions, were widely explored (Wang et al., 2020; Xu et al., 2020). Generally, it was possible to design the expected heterojunctions type considering the conduction bands (CB) and valence bands (VB) positions of both components based on the perceived photoinduced charge transfer pathway (Huang et al., 2017; Zhao et al., 2021a). For example, considering that the CB of NH2-UiO-66 and S-TiO2 were determined as -0.60 eV and -0.35 eV vs. NHE and the VB potentials were calculated as 2.30 eV and 2.84 eV vs. NHE (Li et al., 2020b), it can be expected to construct type II S-TiO2/NH2-UiO-66 heterojunction as illustrated in Fig. 2a, which was confirmed by the determination of the photo-produced active species. The as-prepared S-TiO2/NH2-UiO-66 can photocatalytically reduce the Cr(VI) into Cr(III) primarily by the photo-yielded electrons accumulated on the CB of S-TiO2, and secondarily by  $\cdot O_2^-$  radicals. And the S-TiO<sub>2</sub>/NH<sub>2</sub>-UiO-66 can also oxidate bisphenol A (BPA) into small molecules with the aid of holes accumulated on the HOMO of NH<sub>2</sub>-UiO-66 and the formed  $\cdot O_2^-$  radicals on the CB of S-TiO<sub>2</sub> (Fig. 2a).

Things are not always what they happened. For example, the type II

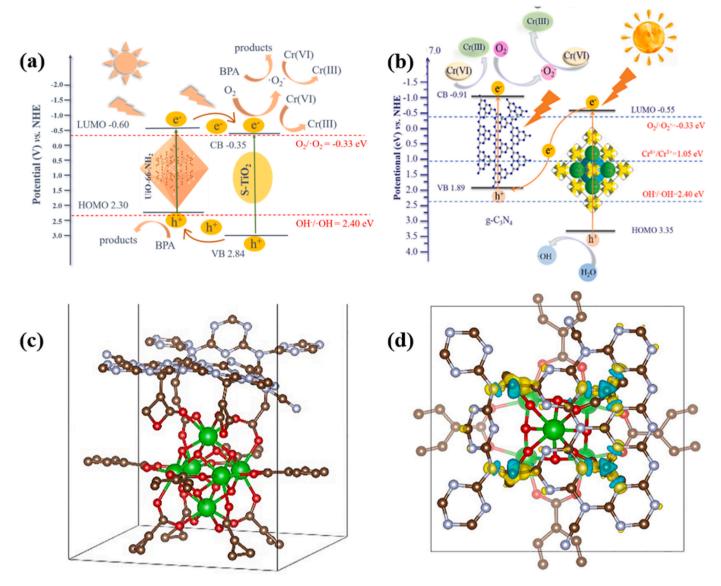


Fig. 2. (a) The photocatalysis mechanism for Cr(VI) reduction and BPA degradation over type II S– $TiO_2/UiO$ -66-NH<sub>2</sub> heterojunction (Li et al., 2020b); (b) The mechanism of the photocatalytic Cr(VI) reduction over Z-scheme g- $C_3N_4/UiO$ -66 heterojunction (Yi et al., 2019). (c) The most stable structure of g- $C_3N_4$  on UiO-66 surface after geometry relaxation (The green, brown, light blue and red spheres represent Zr, C, N and O atoms, respectively). (d) The electron density difference of carbon atoms UiO-66 and g- $C_3N_4$  in g- $C_3N_4/UiO$ -66 heterojunction from the top view (The yellow and the blue regions represent increasing and decreasing of the electron density, respectively) (Yi et al., 2019). (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)

composites can be expected to form between UiO-66 and  $g\text{-}C_3N_4$  based on their HOMO/LUMO and CB/VB positions (Fig. 2b). However, some solid evidence (like XPS determination, active species identification) affirmed that the formed heterojunction is Z-scheme rather than the expected type II. Also, the DFT calculation was adopted to further clarify the flow direction of the photo-induced electrons. It was observed that in the as-prepared  $g\text{-}C_3N_4$ /UiO-66, the electron density around carbon atoms on UiO-66 was lower than that of the pristine UiO-66 (Fig. 2c and d), and the electron density around carbon and nitrogen atoms of the  $g\text{-}C_3N_4$  was higher than those of the individual  $g\text{-}C_3N_4$  (Yi et al., 2019). As well, the DFT calculation was affirmed as a powerful tool to determine the heterojunction types in other works (Wei et al., 2021; Zhao et al., 2021b). It was believed that the Z-scheme composite can achieve the photoinduced charge carriers with stronger redox ability on two components to boost the photocatalysis performance (Huang et al., 2017).

Besides the above-mentioned proofs like XPS determination and active species identification, the combination of DFT calculation and Pt

or PbO2 deposition test can further clarify the charge transfer tracking between the components. It was deemed that Pt<sup>4+</sup> in the H<sub>2</sub>PtCl<sub>6</sub> can be reduced into Pt<sup>0</sup> nanoparticles by the excessive photo-induced electrons, and the Pb<sup>2+</sup> can be oxidized into PbO<sub>2</sub> nanoparticles by the accumulative holes, which can be used to monitor the electron and hole flow (Jiang et al., 2018). It can be observed that the Pt<sup>0</sup> nanoparticles were primarily deposited on the NH2-UiO-66 octahedron in the Bi5O7. I/NH<sub>2</sub>-UiO-66 heterojunction (Fig. 3a) (Zhao et al., 2021b). Detailly, the photo-induced electrons were accumulated on the LUMO of NH2-UiO-66 as the holes on its HOMO were consumed by the electrons excited from the Bi<sub>5</sub>O<sub>7</sub>I via Z-scheme mechanism (Zhao et al., 2021b). The Z-scheme charge carriers transfer process was affirmed by DFT calculation, in which the average 0.2026 electrons were transferred from Bi<sub>5</sub>O<sub>7</sub>I to NH<sub>2</sub>-UiO-66 at their interface (Fig. 3b). In our another case, the photo-deposition of PbO2 NPs over WO3/MIL-100(Fe) resulted from that the accumulative photo-generated holes left over the VB of the WO3 led to the production of PbO<sub>2</sub> nanoparticles from Pb<sup>2+</sup> (Fig. 3c) (Wang et al.,

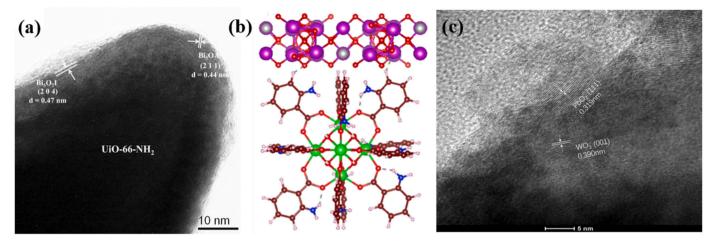


Fig. 3. (a) HRTEM image of photo-deposited Pt nanoparticles over the BU-5. (b) optimized structure model of the Bi<sub>5</sub>O<sub>7</sub>I/UiO-66-NH<sub>2</sub> heterojunction. The brown, blue, red, pink, green, purple, and grey spheres represent the C, N, O, H, Zr, Bi and I atoms, respectively (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.) (Zhao et al., 2021b), (c) HRTEM images of photo-deposited PbO<sub>2</sub> over M80W120 (Wang et al., 2021b).

### 2021b).

Some solid mediators like Au, Ag, and Cu could be introduced to construct the solid mediator Z-scheme photocatalysts for boosting the photocatalytic performances (Lai and Lee, 2021; Li et al., 2016). Our group fabricated a solid mediator Z-scheme Ag/Ag<sub>3</sub>PO<sub>4</sub>/NH<sub>2</sub>-MIL-125 photocatalysts adopting the photo chemical reduction deposition strategy (Zhou et al., 2020b). Under the irradiation of visible light, the formed electrons over the conduction band (CB) of Ag<sub>3</sub>PO<sub>4</sub> can be transported to the Ag<sup>0</sup> nanoparticles (NPs) via the Schottky barrier due to that the Fermi level of Ag<sup>0</sup> NPs is higher than the CB of Ag<sub>3</sub>PO<sub>4</sub>. As well, the produced holes on the highest occupied molecular orbital (HOMO) of NH<sub>2</sub>-MIL-125 may migrate into the Ag<sup>0</sup> NPs, as the HOMO of NH<sub>2</sub>-MIL-125 is more positive than the Fermi level of Ag<sup>0</sup> NPs (Fig. 4a). The proposed photocatalysis mechanism was further affirmed by the photo-deposition Pt<sup>0</sup> element experiment, in which the Pt<sup>4+</sup> in the Cl<sub>6</sub>H<sub>2</sub>Pt was reduced into the Pt<sup>0</sup> by the electrons accumulated on the LUMO of NH<sub>2</sub>-MIL-125 (Fig. 4b and c).

Besides being adopted as solid mediator Z-scheme photocatalysts, the precious metals like Ag and Pd nanoparticles can be coated on the MOFs to accomplish boosted separation of photo-induced charge carriers and enhanced visible-light absorption activity (Shen et al., 2013b; Zhang et al., 2019). Our group deposited Ag nanoparticles on NH<sub>2</sub>-UiO-66 to produce Ag–NH<sub>2</sub>-UiO-66 composites from AgNO<sub>3</sub> and NH<sub>2</sub>-UiO-66 as precursors under high power Xe lamp, which

demonstrated superior Cr(VI) reduction to the pristine NH<sub>2</sub>-UiO-66 due to the improved charge carrier transfer at the interface (Zhou et al., 2020b).

Considering that MOF-derivatives might maintain the advantages of the MOF precursors like adjustable pore and flexible designability (Hang et al., 2022; Li et al., 2021; Zhang et al., 2021; Zhao et al., 2020a), a new approach was proposed to fabricate MOF-derivative/MOF composites like  $\rm In_2S_3@NH_2\text{-}MIL\text{-}68(In)@In_2S_3$  (Hou et al., 2020) and NH2-MIL-125@TiO2 (Zhang et al., 2018b) heterojunctions. Our group obtained marigold-flower-like TiO2/MIL-125 composite (Fig. 5) adopting in-situ post-solvothermal treatment strategy to partially derivate MIL-125 with the aid of thioacetamide (Li et al., 2021). It was expected that the core-shell TiO2/MIL-125 composite as type-II photocatalyst exhibited better photocatalytic Cr(VI) reduction activity than both the completely derived TiO2 from MIL-125 and the pristine MIL-125.

### 2.3. The MOFs composites for simultaneously photocatalytic Cr(VI) reduction and adsorptive Cr(III) removal

It was well known that the accumulative formed Cr(III) in the system might inhibit the forward reaction from Cr(VI) to Cr(III) and result into masking the active sites by the potentially formed Cr(OH)<sub>3</sub> precipitates. Different from the expectation of red-shift of exciting light via introducing different visible light responsive semiconductor photocatalysts,

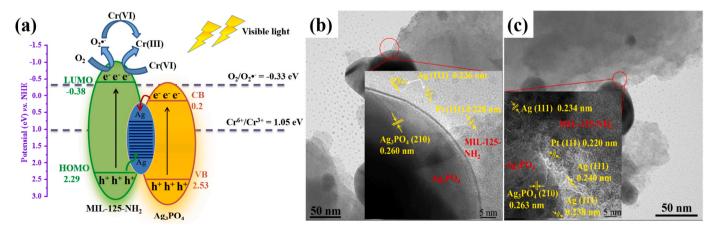


Fig. 4. (a) Schematic illustration and the proposed mechanism of photocatalytic Cr(VI) cleanup over AAMN-120 under visible light irradiation, (b) and (c) HRTEM image of Pt-loaded AAMN-120 (inserted figure is HRTEM of corresponding TEM area) (Zhou et al., 2020b).

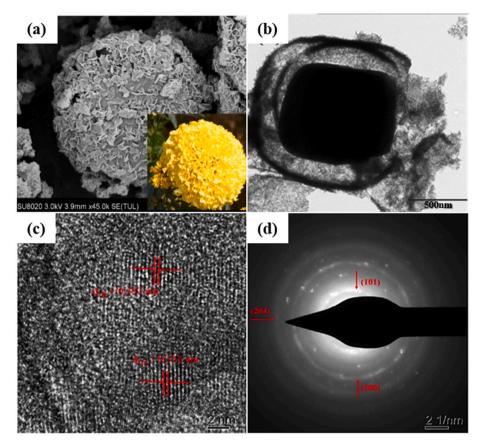


Fig. 5. (a) SEM images of TiO<sub>2</sub>/MIL-125 (MT) and Inca marigold (inset), (b) TEM image of MIL-125, (c) HRTEM and (d) SAED of TiO<sub>2</sub> shell in MT (Li et al., 2021).

our group selected titanate nanotubes (TNTs), an effective and emerging adsorbent toward different cationic heavy metals (Liu et al., 2013, 2014, 2016), to fabricate TNT/BUC-21 composite for the simultaneous photocatalytic Cr(VI) reduction and adsorptive Cr(III) removal (Wang et al., 2019). The introduced TNTs could effectively capture the formed Cr(III) ions to push the forward reaction of Cr(VI) reduction and to avoid the formation of Cr(OH)<sub>3</sub> precipitates (Fig. 6a). It was observed that the TNTs can boost the Cr(VI) reduction (Fig. 6b and c) and total Cr removal (Fig. 6d) performances due to that TNT can adsorb the produced Cr(III) via both the electrostatic attractions and ion exchange (Fig. 6a). BUC-21 can reduce Cr(VI) and TNT can adsorb the formed Cr(III) under higher pH (like 5.0–9.0), which achieved perfect match of BUC-21 and TNT to construct bifunctional TNT/BUC-21 composite for simultaneously photocatalytic Cr(VI) reduction and adsorptive Cr(III) removal.

## 3. From powder MOFs for batch experiment to immobilized ones for continuous operation

The powder photocatalysts can offer sufficient contact interface and facilitate mass transport process of the wastewater pollutants (Liu et al., 2021). However, they are suffering from the difficult recovery/recycling in aqueous solution (Du et al., 2022). Some special reactors applicable to powder photocatalysts were designed and adopted to conduct photocatalysis process, however, massive energy should be input to press the water molecules through the membrane and retain the photocatalyst particles in the reactor (Wang et al., 2021a). To overcome the disadvantages like difficult recovery and reuse, it was feasible to immobilize MOF photocatalysts on some substrates like sand particles (Sadeghian et al., 2021), cellulosic foam (Liu et al., 2021), polymer composite bead (Valizadeh et al., 2020), and  $\alpha$ -Al<sub>2</sub>O<sub>3</sub> sheet (Du et al., 2019; Zhao et al., 2022).

Our group prepared the NH<sub>2</sub>-UiO-66(Zr/Hf) membranes on the

 $\alpha$ -Al<sub>2</sub>O<sub>3</sub> sheet using a reactive seeding strategy, which were adopted to accomplish photocatalytic Cr(VI) reduction under the illumination of sunlight (Du et al., 2019). The results revealed that the as-prepared NH<sub>2</sub>-UiO-66(Zr) and NH<sub>2</sub>-UiO-66(Hf) membranes displayed Cr(VI) reduction (initial concentration being 5.0 mg/L) efficiencies of >98.0% within 120 min. Moreover, the photocatalytic Cr(VI) reduction efficiencies of the NH<sub>2</sub>-UiO-66(Zr) membrane maintained >94.0% during the continuous 20 runs' operation up to 2400.0 min (Fig. 7a). Even when the photocatalytic Cr(VI) reduction tests were carried out in the simulated wastewater containing Cr(VI) under the real sunlight, >93.0% reduction efficiency was accomplished, indicating the potential application for the real wastewater under the real sunlight. As well, the as-prepared NH<sub>2</sub>-UiO-66(Zr) membrane demonstrated good water stability during the long-term operation, which was affirmed by the perfect octahedron morphology of the used NH<sub>2</sub>-UiO-66(Zr) for 20 runs (Fig. 7b).

Inspired by the successfully photocatalytic Cr(VI) reduction of the NH2-UiO-66(Zr) membrane, our group continued to immobilize NH2-MIL-101(Fe) onto  $\alpha$ -Al<sub>2</sub>O<sub>3</sub> sheet to construct NH<sub>2</sub>-MIL-101(Fe)@Al<sub>2</sub>O<sub>3</sub> (MA) (Zhao et al., 2022). Different from the previous NH<sub>2</sub>-UiO-66(Zr) membrane on Al<sub>2</sub>O<sub>3</sub> sheet, it was difficult to observe the clear morphology of NH<sub>2</sub>-MIL-101(Fe) in MA (Zhao et al., 2022). Therefore, the SEM-mapping, HRTEM-mapping, AFM and XPS analysis associated with argon ion etching were adopted to affirm the immobilization of NH<sub>2</sub>-MIL-101(Fe) in MA. The as-prepared MA demonstrated outstanding photocatalytic Cr(VI) reduction performance adopting oxalic acid as hole consumer under white light irradiation, in which 100% Cr(VI) with initial concentration being 5.0 mg/L can be reduced into Cr(III) for 20 runs' operation within 400 min (20 min each run), as displayed in Fig. 8a. The continuous Cr(VI) reduction was also tested, and the results revealed that 100% Cr(VI) with initial concentration of 5 mg/L can be eliminated under white light up to 30 h (Fig. 8b). Also, the

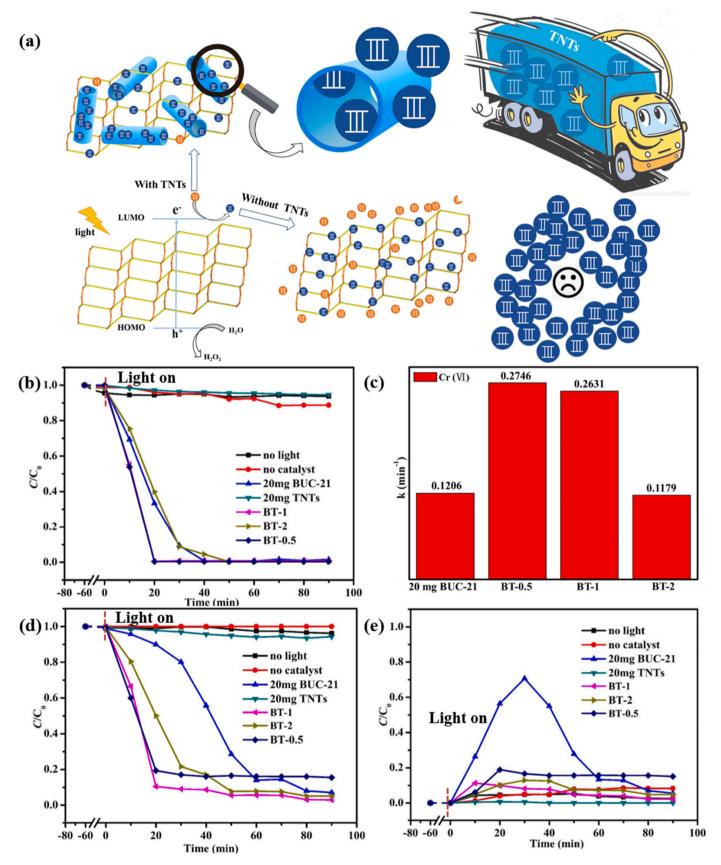


Fig. 6. (a) The simultaneous photocatalytic Cr(VI) reduction and adsorptive Cr(III) removal mechanisms over TNT/BUC-21 composite, (b) efficiencies of photocatalytic Cr(VI) reduction, (c) the corresponding reaction rate constants (k) and (d) Cr(VI) removal efficiencies of control experiments and with different photocatalysts (Wang et al., 2019).

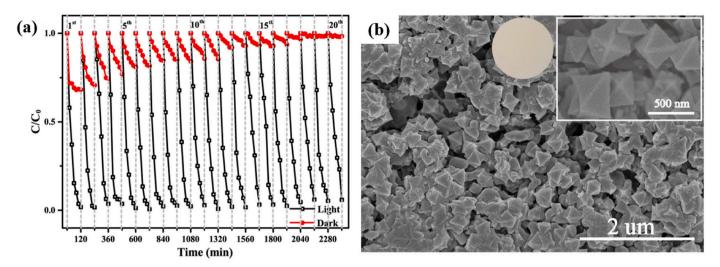


Fig. 7. (a) The long term (20 successive cycles for 2400 min) photocatalytic Cr(VI) reduction performances of the  $NH_2$ -UiO-66(Zr) membrane under the irradiation of the simulated sunlight. (Note: the red line in the figure is for the adsorption activities of the as-prepared membrane); (b) the octahedron morphology of the  $NH_2$ -UiO-66(Zr) in the used membrane after 20 runs' operation (Du et al., 2019). (For interpretation of the references to color in this figure legend, the reader is referred to the Web version of this article.)

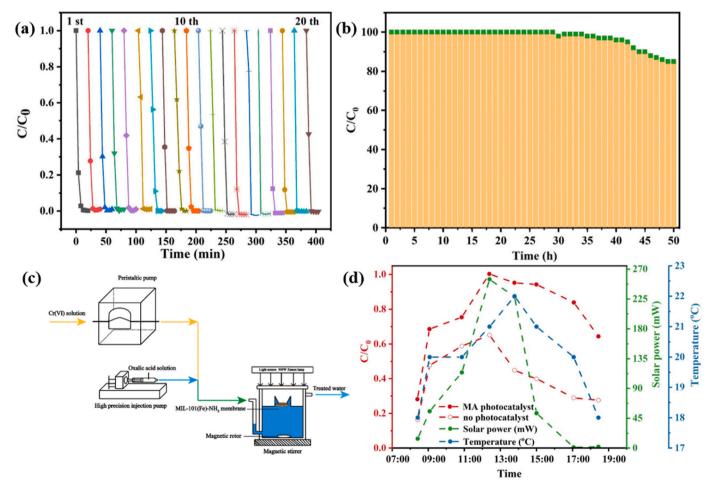


Fig. 8. (a) The long term (20 successive cycles for 400 min) photocatalytic Cr(VI) reduction performances of MA under the irradiation of white light in batch-scale, (b) the continuous Cr(VI) reduction performance of MA in fixed bed reactor, (c) the illustration of the fixed bed reactor for the continuous Cr(VI) reduction test and (d) the continuous Cr(VI) reduction under real sunlight from 8:00 a.m. to 6:00 p.m. Reaction conditions: the actual MIL-101(Fe)–NH<sub>2</sub> being ca. 15.0 mg per sheet, Cr (VI) = 5 mg/L, oxalic acid = 0.8 mM (batch test) and 20 mM (continuous test), 300 W white light irradiation, initial solution pH = 3.4 (Zhao et al., 2022).

continuous Cr(VI) reduction was achieved under real solar light, in which 100% Cr(VI) reduction efficiency was accomplished as the optical power reaching 280.3 mW at 12:00 (Fig. 8c and d). This work implied that the continuous operation adopting MA as photocatalyst in self-developed fixed bed reactor with oxalic acid as hole capturer under both white light and real solar light was feasible and promising.

### 4. The influence factors of the environmental applications

In recent years, increasing researchers focused on the potentially practical applications of photocatalytic Cr(VI) reduction (Ge et al., 2021; Zhang et al., 2017), in which it was essential to consider the influence of the pH, co-existing inorganic ions and organic matters (Hu et al., 2019).

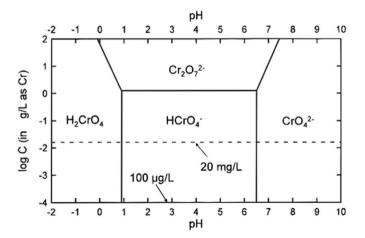
### 4.1. The influence of pH

It was well known that pH might exert noticeable influence toward the surface charge (Zeta potential) of the photocatalysts (Ji et al., 2022) and the chemical species of Cr(VI) (Zhang et al., 2022), which can further alter the adsorptive interactions and the photocatalytic Cr(VI) reduction mechanism. Under the different pH conditions, the Cr(VI) displays various forms like  $CrO_4^{2-}$ ,  $Cr_2O_7^{2-}$ ,  $H_2CrO_4$  and  $HCrO_4^{-}$  (Fig. 9) (Li et al., 2021c; Zhao et al., 1998). As 2.0 < pH < 6.0, Cr(VI) exists primarily as HCrO<sub>4</sub> and Cr<sub>2</sub>O<sub>7</sub><sup>2</sup>. The photocatalytic Cr(VI) reduction reactions followed the Eqs. (1) and (2) under acidic conditions, in which the excessive H<sup>+</sup> might facilitate the reduction from Cr(VI) to Cr(III) (Yi et al., 2019; Zhao et al., 2020b; Zhou et al., 2020b). While, as pH > 7.0, the  $CrO_4^{2-}$  is the primary specie, leading to a reduction reaction as expressed in Eq. (3) (Li et al., 2020a), in which the Cr(OH)<sub>3</sub> precipitate might be formed to mask the active sites over the photocatalysts. Thus, it was preferred that the photocatalytic Cr(VI) reduction was carried out under acidic conditions. However, the acidic reaction conditions required the photocatalysts to possess the property of acid-proof. Some MOFs like UiOs and MILs constructed from hard metal ions like Zr<sup>4+</sup>, Ti<sup>4+</sup>, and Fe<sup>3+</sup> and organic carboxylate ligands are stable in the acid environment, which can be used as photocatalysts to conduct Cr(VI) reduction following the Eqs. (1) and (2). However, some MOFs composed of soft metal ions like Zn<sup>2+</sup> and Co<sup>2+</sup> and imidazolate were not suitable to reduce Cr(VI) to Cr(III) under acidic condition due to that they can be destroyed.

$$Cr_2O_7^{2-} + 14H^+ + 6e^- \rightarrow 2Cr^{3+} + 7H_2O$$
 (1)

$$HCrO_4^- + 7H^+ + 3e^- \rightarrow Cr^{3+} + 4H_2O$$
 (2)

$$CrO_4^{2-} + 4H_2O + 3e^- \rightarrow Cr(OH)_3 + 5OH^-$$
 (3)



**Fig. 9.** Relative distribution of different Cr(VI) species in water as a function of pH and total Cr(VI) concentration (Zhao et al., 1998).

### 4.2. The influence of co-existing matters

The co-existing matters might influence the adsorption and reduction activities of the photocatalysts toward Cr(VI) (Yao et al., 2014). Generally, the photocatalysts with positive surface charge are inclined to attract the negative Cr(VI) ions to their active sites. The co-existing anions like Cl $^-$ , NO $_3^-$ , SO $_4^2^-$ , CO $_3^2^-$ , HCO $_3^-$ , HnPO $_4^{n-3}$  might compete with Cr(VI) to occupy the active sites. While, it was deemed that the co-existing cations like K $^+$ , Na $^+$ , Ca $^{2+}$  and Mg $^{2+}$  with stable electronic structures exerted negligible influence to the photocatalysis process considering that they would not to consume the photo-produced electrons or holes resulting from the stable and highest oxidation states (Li et al., 2022).

The Box-Behnken experimental design methodology was introduced to further clarify the influence of the co-existing matters on the Cr(VI) reduction process. In the Cr(VI) reduction system adopting BUC-21/Bi<sub>24</sub>O<sub>31</sub>Br<sub>10</sub> composite as photocatalyst (Zhao et al., 2020a), four influence variables like Cl $^-$ , NO $_3^-$ , SO $_4^{2-}$  and DOM (dissolved organic matters) were selected to explore the influence toward photocatalysis activity. The ANOVA results revealed that the influence of the four variables on Cr(VI) reduction followed the order of DOM > NO $_3^-$  > Cl $^-$  > SO $_4^{2-}$ . The optimum efficiency fitted by the Box-Behnken methodology demonstrated that the preferred concentrations of Cl $^-$ , NO $_3^-$ , SO $_4^{2-}$  and DOM were 36.4, 9.5, 76.3 and 10.5 mg/L, respectively, which will provide useful information to adjust parameters for large-scale and practical Cr(VI) reduction application.

The influences of the co-existing organic matters varied under different photocatalysis systems. It was deemed that the organics acting as hole scavenger might consume the photo-induced holes, effectively inhibit the recombination of holes and electrons and supply more electron to reduce Cr(VI). Our group investigated the roles of tartaric acid > citric acid > oxalic acid toward the Cr(VI) performance of BUC-21/g-C $_3N_4$  composite photocatalyst, indicating that tartaric acid with two  $\alpha$ -hydroxyl groups was inclined to consume the holes to achieve best Cr (VI) reduction performance (Gao et al., 2019). However, if the Cr(VI) reduction was originated from the superoxide radicals and electrons, the co-existing organics might inhibit the photocatalytic Cr(VI) reduction activity due to that the superoxide radicals might participate both the Cr (VI) reduction and organic oxidation (Li et al., 2020a; Wang et al., 2017).

In some photocatalysis systems like (OH)<sub>2</sub>-UiO-66-20% photocatalyst (Li et al., 2021b), the introduced organics like tartaric acid can achieve "one stone killing three birds", i.e. consuming the photo-yielded electrons for enhanced separation of electron-hole pairs, reacting with photocatalyst to form charge-transfer-complex for boost the electron transfer, as well as producing COO· radicals to directly participate the Cr(VI) reduction.

### 4.3. The influence of light

The light with different light intensities and wavelengths will provide different number of photons, which heavily influence the Cr(VI) reduction efficiencies. Generally, the increasing light intensity resulted into the increasing photons absorbed by the photocatalyst and the enhancing photocatalysis efficiency (Liu et al., 2009). Our group observed that both the Cr(VI) reduction efficiencies (40.7%, 76.7% and 99.9%) and rates (expressed as the pseudo-first-order kinetic constant of Cr(VI) reduction, 0.00492  $\pm$  0.00027, 0.01381  $\pm$  0.00092, and 0.06287

 $\pm$  0.00314 min<sup>-1</sup>) increased with the stronger white light intensities (15.18, 24.25 and 46.07 mW/cm<sup>2</sup>) (Zhao et al., 2020a). The apparent quantum efficiency (AQE), as a critical index to assess the photocatalysis performance, was introduced to test the influence of light with different wavelengths toward Cr(VI) reduction following Eq (4). Our group investigated the AQEs of S–TiO<sub>2</sub>/NH<sub>2</sub>-UiO-66 under the irradiation of

light with different wavelengths. The calculated AQEs were 0.56%, 0.48%, 0.33%, 0.05% and 0.04% at 315 nm, 330 nm, 365 nm, 380 nm, 400 nm and 420 nm, respectively, which corresponded perfectly to the plot trend of UV–vis DRS spectrum of S–TiO $_2$ /NH $_2$ -UiO-66. The investigation of the influences of light intensity and wavelength of the photocatalyst toward the Cr(VI) reduction can provide insight to the large-scale operation under special light or real sunlight (Jain et al., 2021).

$$AQE(Cr) = \frac{3 \times [number \ of \ reduced \ Cr(VI)] \times hc}{Pt\lambda}$$
(4)

Where, P is the optical power, t is the irradiation time,  $\lambda$  is wavelength of light, and h and c are Planck constant (h = 6.62607015  $\times$  10<sup>-34</sup> J s) and light speed (299,792,458 m s<sup>-1</sup>), respectively.

#### 5. Conclusion and outlooks

We once proposed some challenges (i.e. how to improve the stability and electronic conductivity) in the previous review paper concerning Cr (VI) reduction adopting MOFs-based materials as photocatalysts. In the past five years, our research findings revealed that it was a feasible and effective strategy to improve the stability via immobilizing the MOFs on the substrate, which can accomplish continuous and long-term operation. In the future, more efforts should be put to immobilize uniform and nano-sized MOFs on the different large-sized substrates, considering the substrate-MOF interactions. Up to now, increasing findings demonstrated that the introduction of secondary semiconductor or even conductor to fabricate the composites with MOFs was successful to overcome the shortcoming of poor electronic conductivity. Especially, some organic conductor polymers like polyaniline (PANI) were adopted to not only enhance the electronic conductivity but also facilitate the shaping (film preparation, electrospinning).

Generally, Cr(VI) can be reduced into Cr(III) directly by photo-induced electrons, in which the superoxide radicals originated from the interaction between photo-yielded electron and dissolved oxygen might also sideways participate in reduction process. The introduction of organics to the photocatalytic Cr(VI) system might consume the formed holes to inhibit the re-combination of the charge carries. Especially, besides as hole scavenger and reductant, some co-existing organics like tartaric acid, citric acid and oxalic acid can form charge-transfer-complex for electron transfer and even produce COO- radicals to improve the Cr(VI) reduction. From this point, it was necessary consider the compositions of the real wastewater containing Cr(VI) to select suitable MOFs photocatalysts for simultaneous removal of Cr(VI) and organics in the future work.

It was ideal to use the abundant solar light as light source to achieve effective photocatalysis. Some approaches and efforts were used to develop visible light responsive photocatalyst. However, the poor light transmittance in the aqueous solution is a big obstacle. Also, it was awkward to observe that both the photocatalysis efficiency and rate decreased significantly once the light source was shifted from the UV light to the visible light (or real solar light). It was worthy to noting that the commercially available UV light even vacuum UV and short-wave UV (UVC) can provide stronger light intensity and shorter light wavelength to achieve more effective and rapid photocatalysis. From this point, it was important to design and manufacture suitable photocatalytic reactors for water treatment, taking account of various emerging photocatalysts and different light source.

### Author contribution statement

Chong-Chen Wang: Conceptualization, Writing – original draft, Funding acquisition, Supervision, Project administration, Writing – review & editing. Xueying Ren: Resources, Writing – original draft. Peng Wang: Writing – review & editing. Cheng Chang: Resources

### **Declaration of competing interest**

The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

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